

Selenium Biotransformations in an Engineered Aquatic Ecosystem for Bioremediation of Agricultural Wastewater via Brine Shrimp Production

Radomir Schmidt,^{○,†} Prapakorn Tantoyotai,[†] Sirine C. Fakra,[‡] Matthew A. Marcus,[‡] Soo In Yang,[§] Ingrid J. Pickering,[§] Gary S. Bañuelos,^{○,||} Krassimira R. Hristova,^{*,○,†,⊥} and John L. Freeman^{*,○,#,∇}

[†]Department of Land, Air and Water Resources, University of California, Davis, California 95616, United States

[‡]Advanced Light Source, Lawrence Berkeley National Laboratory, Berkeley, California 94720, United States

[§]Department of Geological Sciences, University of Saskatchewan, Saskatoon, Saskatchewan S7N 5E2, Canada

^{||}United States Department of Agriculture, Agricultural Research Service, SJVASC, Water Management Research Division, Parlier, California 93648, United States

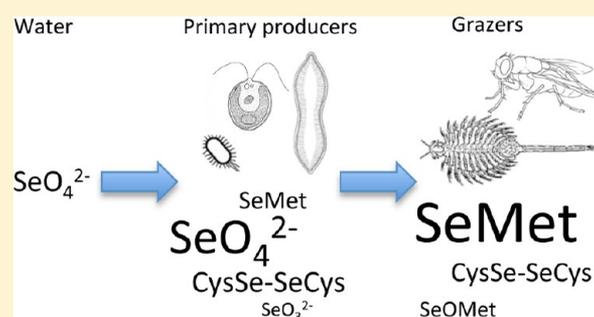
[⊥]Biological Sciences Department, Marquette University, Milwaukee, Wisconsin 53233, United States

[#]Department of Biology, California State University, Fresno, California 93740, United States

[∇]Intrinsyx Technologies Corporation Inc. Space Biosciences Division, NASA Ames Research Center, Moffett Field, California 94035, United States

Supporting Information

ABSTRACT: An engineered aquatic ecosystem was specifically designed to bioremediate selenium (Se), occurring as oxidized inorganic selenate from hypersalinized agricultural drainage water while producing brine shrimp enriched in organic Se and omega-3 and omega-6 fatty acids for use in value added nutraceutical food supplements. Selenate was successfully bioremediated by microalgal metabolism into organic Se (seleno-amino acids) and partially removed via gaseous volatile Se formation. Furthermore, filter-feeding brine shrimp that accumulated this organic Se were removed by net harvest. Thriving in this engineered pond system, brine shrimp (*Artemia franciscana* Kellogg) and brine fly (*Ephydriidae* sp.) have major ecological relevance as important food sources for large populations of waterfowl, breeding, and migratory shore birds. This aquatic ecosystem was an ideal model for study because it mimics trophic interactions in a Se polluted wetland. Inorganic selenate in drainage water was metabolized differently in microalgae, bacteria, and diatoms where it was accumulated and reduced into various inorganic forms (selenite, selenide, or elemental Se) or partially incorporated into organic Se mainly as selenomethionine. Brine shrimp and brine fly larva then bioaccumulated Se from ingesting aquatic microorganisms and further metabolized Se predominately into organic Se forms. Importantly, adult brine flies, which hatched from aquatic larva, bioaccumulated the highest Se concentrations of all organisms tested.



INTRODUCTION

Selenium is a naturally occurring trace element mainly found in soils derived from marine sediments, including Cretaceous shale deposits in western North America.^{1,2} Selenium biochemical properties lead to a narrow boundary between deficiency and toxicity,³ a phenomenon further complicated by a large variety of Se chemical forms with differing levels of biological activity. Human Se deficiency causes multiple diseases, and it is estimated that more than a billion people worldwide ingest insufficient amounts of Se in their diet (less than $10 \mu\text{g Se day}^{-1}$).^{4,5} In central California, marine, pyritic sedimentary rocks in the Coast Ranges are a major source of soluble selenate, sodium chloride, and magnesium and calcium sulfate salts that can cause hyper-salinization of the Western

San Joaquin valley (WSJV) soils.⁶ Water-soluble inorganic selenate (SeO_4^{2-}) is often leached from naturally enriched soils by irrigation and winter rains into surface runoff or effluent waters, accumulated by catch basins or reservoirs and in nearby wetland marshes where Se is evapo-concentrated during the summer.⁶ Phytomanagement is a strategy for mitigating soluble selenate in soil from entering waters of the WSJV that utilizes *Brassica* plants (mustard or canola) to accumulate soluble Se from soils.⁷⁻⁹ Seeds from these plants are in turn used to

Received: January 18, 2013

Revised: April 19, 2013

Accepted: April 26, 2013

produce biodiesel fuel and Se-enriched seed meals are used to supplement livestock animal feed.^{7–9} The increased accumulation of naturally occurring salts, boron (B), and SeO_4^{2-} has worsened in some agricultural areas due to limited freshwater supplies, low winter rainfall, and drought conditions, which reduce the leaching of salts away from the root zone. In Se polluted aquatic environments, such as Kesterson Wildlife Reservoir in WSJV, Se is present primarily as water-soluble selenate (SeO_4^{2-}) $\geq 98\%$ and $\sim 2\%$ selenite (SeO_3^{2-}).² Due to its relative abundance in these environments, ecotoxicity concerns center around SeO_4^{2-} and its biomagnification up the aquatic food chain.^{10,11} Selenium evapoconcentration in aquatic ecosystems has been documented to cause deformity and mortality in fish, waterfowl, shore birds, and mammals.^{12–15}

The food chain transfer and bioaccumulation of organic Se represents another exposure route in addition to the direct uptake of SeO_4^{2-} .^{11,12} Possible biochemical transformations, variability in food chain accumulation, and a strong dependence on the physical site characteristics make it difficult to predict Se pollution risks based on waterborne, inorganic Se concentrations alone.^{11,12} In addition to accumulating Se in biomass, aquatic primary producers are main drivers for Se volatilization in aquatic ecosystems via production of methylated selenides, including dimethylselenide (DMSe) and dimethyldiselenide (DMDSe). These methylated selenides exit the water column into the atmosphere or can be oxidized back to SeO_3^{2-} .¹⁶

The engineered aquatic ecosystem and the drainage waters at Red Rock Ranch (RRR) mimic the conditions present in brackish saline marshes, which are often impacted by selenium, when near low lying agricultural areas adjacent shale derived Se enriched soils. Cyanobacteria, diatoms, and microalgae are the primary producers in this open-air pond system and are the main food source for both brine shrimp and larvae of the brine fly. These same brine shrimp and brine fly are also the two major groups of invertebrates that flourish in highly saline environments and thus play a major role in associated aquatic food webs.^{17,18} *Artemia franciscana* Kellogg (brine shrimp) are Brachiopods. *Ephydra cinerea* (brine fly) larvae and adults occupy diverse habitats such as salt pools, alkaline lakes, and marshes and both typically graze mostly on the algal and bacterial communities. The RRR pond is completely enclosed by netting to prevent incidental exposure of birds.

As part of a larger integrated on-farm drainage management system (IFDM) used to manage excess saline irrigation water, we studied the bioremediation of SeO_4^{2-} laden wastewater in the RRR pond by analyzing the internal Se concentrations of the various food web organisms present in the engineered aquatic ecosystem pond designed to produce brine shrimp enriched in organic Se and omega-3 and omega-6 fatty acids for use in value added nutraceutical food supplements (Supporting Information Figures 1 and 2).

Furthermore, the biological fate of Se was determined with high resolution in mixed bacteria, diatoms, *Picocystis* sp. (microalgae), and *Cladophora* (macro-algae), and the macro-invertebrates brine fly and brine shrimp. We achieved this by directly collecting specimens (brine shrimp and brine fly) or isolating individual microscopic strains and microorganism groups using a variety of controlled growth and culture systems. These sample isolation methods were then combined with the use of advanced analytical techniques including strong anion exchange, high performance liquid chromatography (HPLC), inductively coupled plasma mass spectrometry (SAX-HPLC-

ICPMS), X-ray absorption near edge structure (bulk XANES), micro-focused X-ray fluorescence (μ XRF), and micro-X-ray absorption near edge structure (μ XANES) (Supporting Information Table 1). The total Se accumulation, chemical Se speciation, and Se atom mapping results yielded high molecular resolution into the accumulation and biotransformation of inorganic Se into organic Se in addition to the localization of Se inside aquatic organisms living in the engineered aquatic remediation system.

■ EXPERIMENTAL SECTION

System Design, Location, and Water Chemistry Parameters. At RRR near Five Points, California, the IFDM water has been used in several irrigation cycles by irrigating progressively more salt tolerant crops (Supporting Information Figure 2). The reduced volume of drainage water was collected in the final sump system and the resulting drainage water used in the racetrack pond (Supporting Information Figure 1). The agricultural drainage water was very high in dissolved salts and metals: Se ($6\text{--}12\text{ mg L}^{-1}$), SO_4^{2-} (approximately 65 g L^{-1}), Na^+ (17 g L^{-1}), Mg^{2+} (1 g L^{-1}), Ca^{2+} (0.5 g L^{-1}), K^+ (51 mg L^{-1}), and B (142 mg L^{-1}). The water also typically contained approximately $300\text{ mg NO}_3^{2-}\text{ L}^{-1}$, but was low in NH_4^+ ($1\text{--}3\text{ mg L}^{-1}$), Fe (0.5 mg L^{-1}), P (0.8 mg L^{-1}), Si (0.9 mg L^{-1}), and V (0.01 mg L^{-1}). Other dissolved metals were below their respective detection limits: Zn, Mn, Cd, Cr, Ni, and Al ($<0.05\text{ mg L}^{-1}$); Cu and Pb ($<0.1\text{ mg L}^{-1}$). Sequential application of the drainage water reused 90% of the wastewater produced at RRR, and there was no disposal into rivers or evaporation basins.

Sample Growth and Preparation. Microorganisms. Bacteria and Bacteria Plus Diatom Growth. Mixed bacterial culture medium consisted of $0.2\text{ }\mu\text{m}$ filter (Millipore) sterilized RRR drainage water at approximately 80 g L^{-1} salinity equivalent amended with Na_2SeO_4 to 40 mg Se L^{-1} , 1% v/v glycerol (Fischer), 2% w/v tryptone (Difco), 20 mg L^{-1} Fe-gluconate (Sigma-Aldrich), 6 mg L^{-1} NaH_2PO_4 (Sigma-Aldrich). Initial tests indicated cultures of bacteria, diatoms, and microalgae at ambient Se concentrations did not produce sufficiently high internal Se concentrations for chemical speciation via X-ray absorption near edge spectroscopy (XANES). For bacteria-only culture, 10 mL RRR drainage water filtered through $5\text{ }\mu\text{m}$ sterile syringe filter (Millipore) was added to 90 mL of media. Cultures were incubated overnight at $25\text{ }^\circ\text{C}$ in the dark on an Orbit shaker-incubator (Lab-line) at 100 rpm.

Microalgae Growth. For culturing *Picocystis* sp., the medium consisted of RRR agricultural drainage water at approximately 80 g L^{-1} salinity equivalent amended with Na_2SeO_4 to 40 mg Se L^{-1} . 500 mL culture flasks containing 100 mL culture media were incubated for two weeks under 30W compact fluorescent lights on a MaxQ2000 orbital shaker (Thermo Scientific) at 90 rpm.

Diatom Growth. A mixed diatom culture was isolated by a modification of the protocols of Bruckner and Kroth.¹⁹ Diatom medium consisted of RRR agricultural drainage water at approximately 80 g L^{-1} salinity equivalent amended with Na_2SeO_4 to 40 mg Se L^{-1} . Chloramphenicol $27\text{ }\mu\text{g mL}^{-1}$, streptomycin at $135\text{ }\mu\text{g mL}^{-1}$, and ampicillin at $270\text{ }\mu\text{g mL}^{-1}$ were used as selective agents. The diatom inoculum was prepared by filtering 100 mL RRR water through a $2\text{ }\mu\text{m}$ filter paper. The filter paper was washed with 100 mL 80 g sulfate L^{-1} buffer. The washed filter was placed in 10 mL diatom

medium and gently vortexed for 30 s; 5 mL inoculum was used per 100 mL solution. Flasks were incubated for two weeks under 30 W compact fluorescent lights on a MaxQ2000 orbital shaker (Thermo Scientific) at 90 rpm.

Microbial Sample Harvest. Samples (bacteria, microalgae, and diatoms) were harvested by centrifugation at 5856 g for 20 min at 23 °C. Cells were washed three times in 5 mL 80 g L⁻¹ sulfate buffer. Pellets in 2 mL microcentrifuge tubes were flash frozen in liquid nitrogen and stored at -80 °C until analysis.

Sample Harvest and Preparation from RRR Racetrack Pond. Macroalgae. Cladophora filamentous algae were collected in 50 mL Falcon tubes, washed with sterile 80 mg mL⁻¹ Na₂SO₄²⁻ in Büchner filter funnels lined with Whatman No. 1. filter papers and flash frozen in liquid nitrogen.

Macroinvertebrates. Brine shrimp and brine fly larvae and pupa were collected in 1 L plastic water sampling bottles. Adult flies were collected by trapping in empty 50 mL Falcon tubes. All macroinvertebrate samples (except for adult fly) were washed with sterile 80 mg mL⁻¹ Na₂SO₄²⁻ in Büchner filter funnels lined with Whatman No. 1 filter papers and flash frozen in liquid nitrogen. Brine fly adults were directly flash frozen in liquid nitrogen. All samples were stored at -80 °C until analysis.

Selenium Analysis. Brief outlines of the various Se analysis methods employed are provided below. Detailed experimental methodology is provided in the Supporting Information.

Quantification of Total Se. Aquatic organisms were removed from the freezer and were digested with HNO₃, H₂O₂, and HCl.²⁰ Mineral inorganic elements, including Se, were analyzed by an inductively coupled plasma mass spectrometer (Agilent 7500cx, Santa Clara, CA, U.S.A.). The National Institute of Standards and Technology (NIST) wheat flour (SRM 1567; Se content of 1.1 ± 0.2 μg g⁻¹ DM) and two internal soil standards (sediment collected from Kesterson Reservoir, CA, with a total Se content of 7.5 ± 0.25 and 25 ± 0.87 mg kg⁻¹) were used as the Se quality control standards. Recovery rates of Se in standard materials after acid digestion were over 94%.

Chemical Speciation of Soluble Se. SAX-HPLC-ICPMS. Aqueous Se compounds were separated from insoluble Se compounds via methanol/chloroform extraction.²¹ One vial of each duplicate sample set was treated with Protease XIV as a first step in the extraction. One quarter of each extracted aqueous phase and the entire nonaqueous chloroform phase were evaporated with a heating block at 50 °C (~300 min), acid digested, and analyzed for total Se by ICPMS. Waters Sep-Pak Classic C18 cartridges were used for final cleanup of the remaining aqueous concentrates. Purified eluent was transferred into Agilent screw-top glass HPLC vials and frozen until SAX-HPLC-ICPMS analysis. (For nonprotease digested sample results see Supporting Information Table 4).

An Agilent 1200 HPLC separations module equipped with a Hamilton PRPX-100 strong anion exchange analytical column (10 mm particle size; 25 cm length × 4.1 mm internal diameter) and an Agilent 7500 ICPMS was used for Se speciation of aqueous extracts.²² A single analysis (30 μL injection) was conducted for each of the aqueous extracts. Chromatographic separation of all Se compounds was achieved with an isocratic mobile phase of 5 mM ammonium citrate buffer (pH 5.2) with 2% methanol at a flow rate of 1 mL min⁻¹ and ⁷⁸Se was monitored in real time using ICPMS. Retention times of Se containing peaks were compared to authentic standards (Supporting Information).

Se Distribution and Chemical Speciation. μXRF/μXANES. The distribution of Se, Ca, and Zn were determined using micro-focused X-ray fluorescence (μXRF) mapping and the local speciation was determined using micro-Se K-edge X-ray absorption near-edge structure (μXANES) spectroscopy. Washed samples were flash frozen in liquid nitrogen and taken to beamline 10.3.2 of the Advanced Light Source at the Lawrence Berkeley National Laboratory for analyses.²³ Micro-XRF elemental maps were recorded at 13 keV, using a 15 × 6 μm (vertical) beam, a 15 × 15 μm pixel size, and 50-ms dwell time per pixel, except for the brine fly pupa, which was imaged using a 25 × 25 μm pixel size. Se K-edge μXANES spectra were recorded with a seven-element Ge solid-state detector (Canberra). Spectra were deadtime corrected, pre-edge background subtracted, and postedge normalized using standard procedures.^{23,24} Red Se (white line maximum set at 12660 eV) was used to calibrate the spectra. Least squares linear combination fitting (LCF) of the Se XANES spectra was performed in the 12630 to 12850 eV range using a library of nine standard selenocompounds (Supporting Information), as previously described.²⁵ The error on the percentages of species present is estimated to be ±10%. Data processing was performed with LabVIEW.

Bulk XANES. Samples were removed from storage at -80 °C, macerated in liquid nitrogen using a mortar and pestle, packed into 2-mm path length sample cells, and maintained in liquid nitrogen for bulk Se K-edge XANES analysis. Spectra were collected on beamline 9-3 at the Stanford Synchrotron Radiation Light source (SSRL). Experimental spectra were calibrated with respect to the spectrum of hexagonal Se. Background subtraction, normalization calibration, and further data analyses were carried out according to standard procedures using the EXAFSPAK program suite. Experimental spectra were analyzed by least-squares linear combination fitting.²⁶

Total Se Analysis by Spectrofluorimetry. Total Se in water, tissue, and volatile matrices was analyzed using micronitric acid digestion coupled with fluorescence detection.²⁷ Fluorescence intensity was measured using a spectrofluorimeter equipped Genios microplate reader (Phenix Research Products). A standard curve consisted of 9 SeO₂ (Fisher Scientific) solution standards.

RESULTS AND DISCUSSION

Selenium Water Chemistry. Selenium concentrations in the racetrack pond water (Supporting Information Figure 1) ranged from 6 to 12 mg Se L⁻¹, and the form of Se (determined via SAX-HPLC-ICPMS), in RRR racetrack pond was 97 ± 2% selenate (SeO₄²⁻) and 3 ± 2% selenite (SeO₃²⁻). In both freshwater phytoplanktonic and marine algae communities, SeO₄²⁻ is considered to be more ecologically toxic than SeO₃²⁻.^{28,29} In most WSJV soils, Se is predominately present as SeO₄²⁻, and less abundant SeO₃²⁻, due to irrigation water leaching out local SeO₄²⁻-enriched alluvial fan derived soils^{2,30} (see Supporting Information Figure 1). The RRR agricultural drainage waters also have high sulfate concentrations of up to 100 g L⁻¹ and nitrate concentrations of up to 300 mg L⁻¹. Methylation of Se is one of the primary routes of Se removal from aqueous systems through Se-volatilization,^{7,16,31} however, high nitrate concentrations may decrease Se methylation.³²

Mixture of Bacteria and Diatoms. We initiated our analyses on a mixture of two key Se-utilizing groups of primary producers, the bacteria and diatoms. The average total Se for the mixed bacterial diatom sample harvested and rinsed after

four days growth in RRR drainage water was $1 \mu\text{g Se g}^{-1}$ wet weight (ww). Bulk XANES least-squares analysis demonstrated that forms of Se were 26% SeO_4^{2-} , 11% SeO_3^{2-} , 9% elemental (Se^0), and 54% organic Se as C–Se–C forms (Supporting Information Figure 3, Table 2). These Bulk XANES results demonstrated that SeO_4^{2-} in water was actively being reduced by microbes into SeO_3^{2-} and then incorporated into organic C–Se–C forms (e.g., selenocystathionine (SeCyst) and selenomethionine (SeMet), while a portion of SeO_4^{2-} was reduced to elemental Se^0 .

Mixed bacteria and diatoms were then protease-XIV digested and extracted using methanol chloroform water (MCW). In the upper aqueous phase (60% methanol + 40% water), 72% of the total Se was recovered due in part to an initial protease-XIV digestion, which led to an 1.7 fold greater recovery of total Se than in nonprotease digested samples. Aqueous phase soluble Se species in mixed bacteria and diatoms (identified by SAX-HPLC-ICPMS) were 78% SeO_4^{2-} , 3% SeO_3^{2-} , and 18% organic Se (C–Se–C; 14% SeCyst and 4% SeMet) (Supporting Information Figure 4, Table 3).

Micro-XRF elemental mapping of Se present in a thin smear of mixed bacteria and diatoms demonstrated numerous hotspots, highly concentrated with Se (Supporting Information Figure 5). Further examination of Se species in hotspots by μXANES revealed they were entirely elemental Se^0 , likely in precipitates or granules (Supporting Information Table 5). This is likely due to the reduction of SeO_4^{2-} by sulfate reducing bacteria into SeO_3^{2-} and Se^{2-} intermediates before a final reduction step into elemental Se^0 .

Bacteria. Isolation of a bacteria only mixture from the RRR pond allowed the growth and recovery of bacteria from an overnight culture that contained $2.5 \mu\text{g Se g}^{-1}$ ww. In the soluble aqueous phase of the protease-XIV digested sample 84% of total Se was recovered and a 1.2 fold increase in Se recovery resulted from protease treatment. The aqueous Se forms (identified by SAX-HPLC-ICPMS) inside the washed bacterial samples were 75% SeO_4^{2-} , 3% SeO_3^{2-} , 2% SeCys, 16% free C–Se–C forms (SeCyst and SeMet), and 3% SeMet containing proteins (Supporting Information Table 3).

Similar to sulfur (S), bacteria are capable of four possible transformations of Se that include both oxidation and dissimilatory reduction, the reduction and assimilation of Se into amino acids, methylation, and demethylation.³² Bacteria have also been shown to possess both high and low affinity Se transport systems.^{33,34} While RRR pond drainage water contained essentially 95% SeO_4^{2-} and 5% SeO_3^{2-} , these forms made up only 78% of the internal Se concentration in bacteria and the other 21% were organic Se (Supporting Information Table 3). The internal concentration of $2.5 \mu\text{g Se g}^{-1}$ ww was low compared to the external environment (40 mg Se L^{-1}). The lower Se concentrations in these cells may reflect their initial stage of Se uptake or their ability to efflux Se.

Diatoms. Axenic, mixed diatom cultures grew well in the sterilized RRR water, even at the highest Se concentration of 40 mg Se L^{-1} . The average total Se for the diatom sample grown at 40 mg Se L^{-1} was $59.8 \pm 1.9 \mu\text{g Se g}^{-1}$ ww. The percentage of Se recovery in the aqueous phase of the MCW was 18% for the protease-XIV digested sample and protease-XIV treatment increased soluble Se recovery 1.9 fold. In contrast to bacteria, the aqueous Se forms of the mixed diatom sample grown for two weeks contained only 3% SeO_4^{2-} with 24% SeO_3^{2-} , in addition to 6% selenocysteine (SeCys), and 56% organic C–Se–C forms (45% SeCyst and 11% SeMet) as identified by

SAX-HPLC-ICPMS (Supporting Information Table 3, Figure 4). SeO_3^{2-} was the highest percentage of inorganic Se measured in diatom samples and was 8 times higher than SeO_4^{2-} concentrations in diatoms. Similarly, a high 6:1 SeO_3^{2-} to SeO_4^{2-} ratio was observed using μXANES analysis, although no organic Se forms were observed (Supporting Information Table 5). Selenium is an essential element for freshwater and marine diatoms,^{30,35,36} and μXRF mapping of mixed diatoms isolated from the drainage water showed a relatively uniform Se distribution (data not shown). Unlike higher plants, diatoms utilize selenoenzymes and therefore require a high affinity Se uptake mechanism and specific selenocysteine tRNAs.³⁷ Mixed diatoms isolated from the drainage water were more efficient than the mixed bacteria at reducing selenate into organic forms, whereas the bacteria were better at reducing selenate into elemental Se^0 .

Green Microalga. The unicellular green microalga *Picocystis* sp. was originally isolated from RRR water. In hypersaline Mono Lake, CA, which has similar salinity (approximately 81 mg L^{-1}) to RRR pond water and also a significant SO_4^{2-} component (approximately 13 g L^{-1}), a *Picocystis* sp. accounts for 25% to 50% of the primary production and is heavily grazed by endemic brine shrimp.³⁸ At RRR, the average total Se for the *Picocystis* sp. sample was $1.2 \mu\text{g Se g}^{-1}$ ww. Bulk XANES analysis of the RRR *Picocystis* sp. demonstrated that after two weeks of growth no Se remained as SeO_4^{2-} , and only 5% was in the inorganic form SeO_3^{2-} , while greater than 95% of the Se was in organic Se forms 71% C–Se–C and 24% CysSe–SeCys (Supporting Information Figure 3 and Table 2). Protease XIV treatment increased soluble Se recovery by 4.1 fold, suggesting a high proportion of selenoamino-acids in proteins; the total soluble Se recovery was 21%. While SAX-HPLC-ICPMS analysis suggested high proportion of inorganic Se forms (36% SeO_4^{2-} , 11% SeO_3^{2-} , 21% CysSe–SeCys, and 32% SeMet (Supporting Information Table 3)), the bulk XANES analysis provided evidence the majority of the unrecoverable Se was in organic forms while essentially all inorganic Se had likely been recovered.

The *Picocystis* sp. analyzed by μXRF mapping had Se distributed homogeneously in a droplet of cells and lacked the hot spots observed in mixed bacteria and diatoms (Supporting Information Figure 5), while the Se concentrations were not high enough to provide good Se signal-to-noise ratios for μXANES analyses.

The low internal inorganic Se concentrations in *Picocystis* sp. was consistent with other unicellular green algae such as *Chlorella*, where >85% Se was present as organic Se.⁴¹ The dominant organic Se form in *Picocystis* sp. in this study was SeMet, followed by CysSe–SeCys (Supporting Information Table 3). Selenocystathionine is a compound often found in secondary Se accumulator plants; thus, it may play a similar role in the Se resistance of the RRR *Picocystis* sp. In addition, internal *Picocystis* sp. Se concentrations were approximately 20-fold lower than in the water after two weeks growth, suggesting they have an efficient Se-exclusion or Se-efflux mechanism.

Filamentous Macroalgae (*Cladophora*). Bulk XANES spectra of the RRR race track collected *Cladophora* demonstrated Se mainly in the organic forms 39% CysSe–SeCys, 34% SeMet, and 27% methionine selenoxide (SeOMet) (Supporting Information Figure 3 and Table 2). Importantly, SeOMet may be an intermediate molecule in a biochemical pathway protective against cellular damage caused by reactive oxygen species^{39,40} and may play a role in scavenging reactive

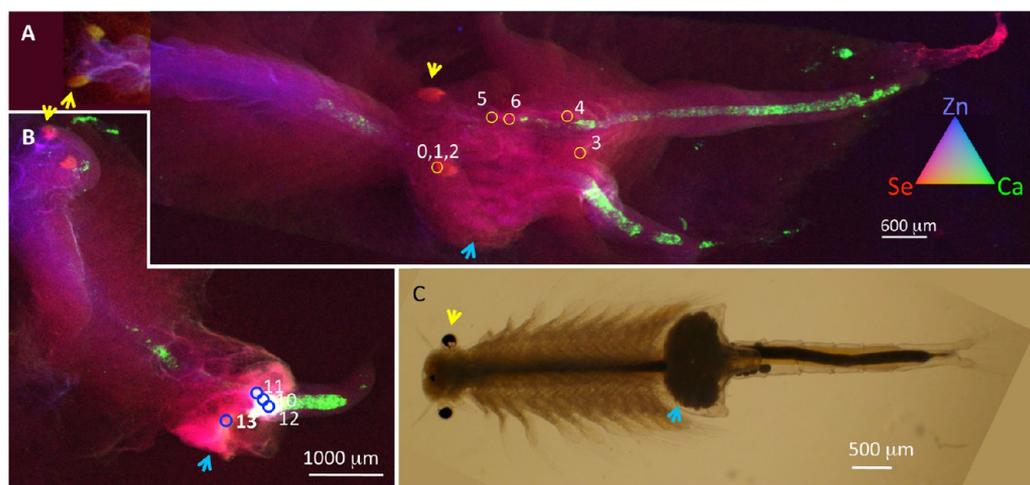


Figure 1. Distribution and speciation of Se in *Artemia franciscana*. μ XRF map showing spatial distribution of Se (coded in red) and other selected elements (calcium in green, zinc in blue) in adult female with eggsac and adult male, coupled (A), in single adult female with eggsac (B). Adult female with eggsac shown for comparison (C). Scale bars = 600 μ m (A), 1 mm (B), 500 μ m (C). Yellow circles in A and B show locations of μ XANES spectra whose fitting results are reported in Supporting Information Table 5.

oxygen species in *Cladophora* or other aquatic organisms. This reaction is shown in Supporting Information Figure 6 with hydroperoxide used as an example. The oxidized selenomethionine (SeOMet) can be reduced by other electron donors, such as diglutathione (GSH-GSH \rightarrow GS-SG) or thioredoxin reductase with NADPH.^{39,40} All bulk XANES samples were prepared in anaerobic conditions (ground and stored in liquid nitrogen) and measured at -263 °C and possible radiation damage during measurement was monitored; thus, the potential oxidation of SeMet by exposure to air or radiation damage during measurement can be disregarded.

Brine Shrimp. *Artemia franciscana* are filter microalgae, diatoms, and bacteria in the water column. The total Se in the RRR brine shrimp sample was $7.2 \mu\text{g Se g}^{-1}$ ww. Bulk XANES analysis of brine shrimp collected from the RRR pond demonstrated that only 3% Se remained in an inorganic form as SeO_3^{2-} , while the majority of Se was found as the organic Se forms 28% CysSe-SeCys, 59% C-Se-C forms (most likely SeMet), and intriguingly 10% SeOMet, similar to the *Cladophora* (Supporting Information Figure 3 and Table 2). This result further suggests SeMet may either be easily oxidized in aquatic systems or could be playing a role in the scavenging of reactive oxygen species inside aquatic organisms. The percentage of Se recovery into the aqueous phase for the protease-XIV digested MCW extracted brine shrimp sample was 63.8% of total Se and protease-XIV treatment increased soluble Se recovery by a notable 3.9 fold.

Micro-XRF mapping revealed that, in adult brine shrimp, the muscle along the back, both compound eyes and cysts (eggs) had accumulated the highest Se concentrations (Figure 1). The forms of Se in adult brine shrimp were different from the Se forms in phytoplankton (bacteria, diatoms, and microalgae) that constitute most of their diet. In particular, brine shrimp had very little inorganic SeO_4^{2-} or SeO_3^{2-} . SAX-HPLC-ICPMS showed that brine shrimp mainly contained organic Se forms occurring as 15% CysSe-SeCys, 81% SeMet, and 2% SeOMet (Supporting Information Figure 4 and Table 3).

Micro-XANES results (Supporting Information Table 5) also supported the bulk XANES and SAX-HPLC-ICPMS data in that at the specific localized Se hot spots, forms were all C-Se-C. These findings suggest that adult brine shrimp efficiently

convert inorganic Se and or accumulate mainly organic Se. The observation of SeMet concentration in brine shrimp eggs (cysts) is of importance to understanding Se in the food web considering these cysts are a major food source for small fish and aquatic insects.

In order to better understand the accumulation of Se in brine shrimp in response to changing water Se concentrations, we analyzed total Se in brine shrimp harvested after controlled laboratory growth under various Se concentrations. Freeze-dried brine shrimp samples analyzed by spectrofluorimetry showed proportional increases in total internal Se concentrations in direct correlation with water Se concentrations (Supporting Information Figure 7). This means that at water concentration of 10 mg Se L^{-1} , harvested brine shrimp contained $\sim 10 \text{ mg Se kg}^{-1}$. At least one study has shown aquatic macroinvertebrates unable to get rid of internal Se once acquired,⁴¹ though more recent results indicate slow Se decrease in macroinvertebrates following dose of a Se pulse.⁴² The internal Se concentration of brine shrimp in this study appeared to mimic water Se concentration and ratios of total Se were similar to those observed at Kesterson reservoir. Interestingly, these Se ratios were much lower than the ratios observed in waters with low SO_4^{2-} content,^{41,43} implicating a synergistic effect of SO_4^{2-} in water on total Se accumulation in aquatic organisms.

Brine Fly (*Ephydra cinerea*). The entire brine fly lifecycle takes place in the water or on the surface of brine lakes and *E. cinerea*, tolerating higher salt concentrations than other Ephydrids, thrive in these environments.⁴⁴ Adult *E. cinerea* deposit sticky eggs on the water surface, which hatch into larvae that feed on epilithic algae, and show a clear preference for diatoms.⁴⁵ The forms of Se in the brine fly (larva and adult) were significantly different from the Se forms in their diet (diatoms, bacteria, and microalgae) (Figure 3, Supporting Information Tables 2 and 3). Bulk XANES analysis of brine fly larva and pupa Se demonstrated that the majority was organic Se occurring as 10% SeCys, 25% CysSe-SeCys, and 52% C-Se-C forms (most likely SeMet). Similar to *Cladophora* and brine shrimp, 13% of Se was SeOMet (Supporting Information Figure 3 and Table 2). The percentage of Se recovery into the aqueous phase for the protease-XIV digested brine fly larva

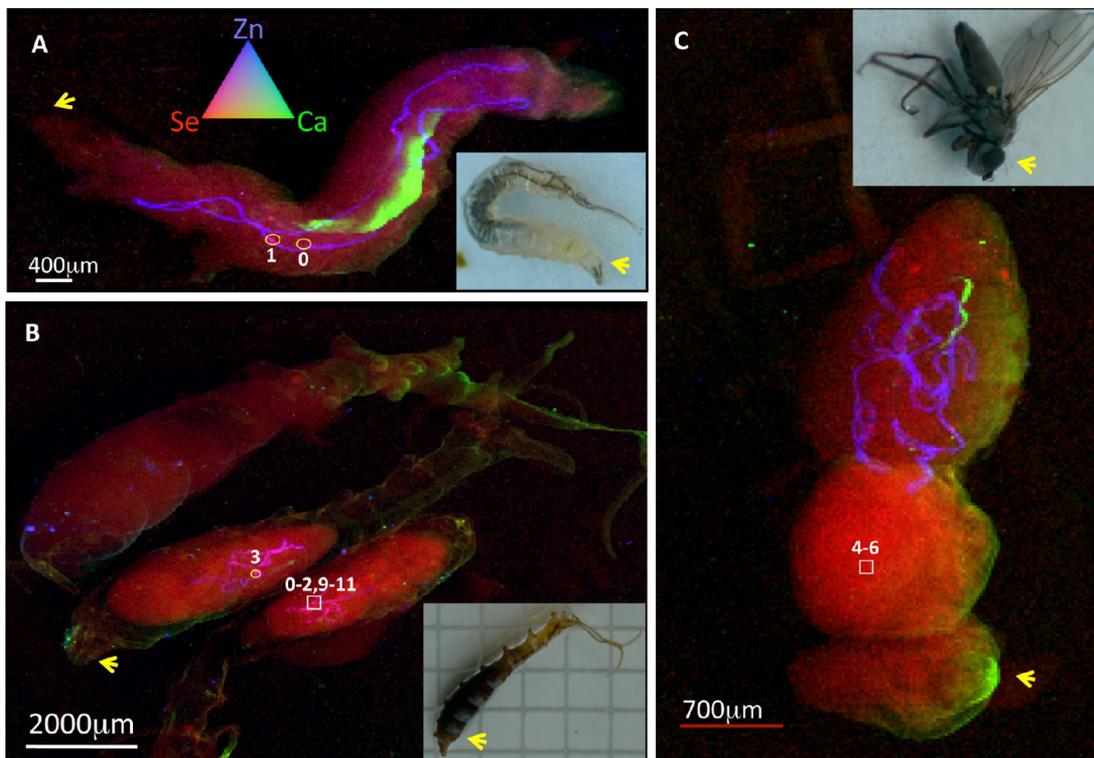


Figure 2. Distribution and speciation of Se in the brine fly *Ephydra cinerea*. μ XRF map showing spatial distribution of Se (coded in red) and other selected elements (calcium in green, zinc in blue) in fly larva (A), three pupae (B), and adult fly (C). Circles in A, B, and C show locations of μ XANES spectra whose fitting results are reported in Supporting Information Table 5. Light micrographs of similar individuals are shown as insets.

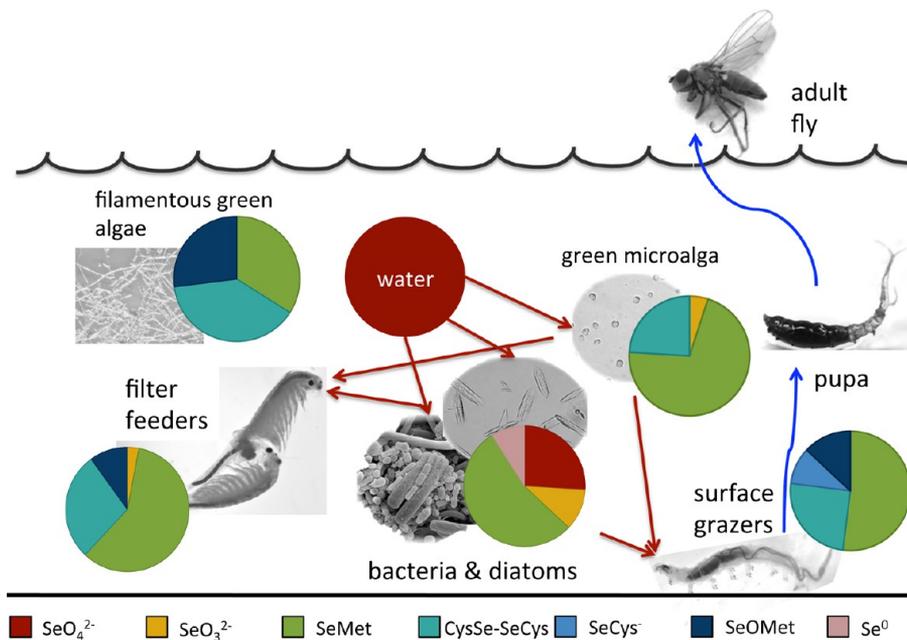


Figure 3. Community Se distribution—conceptual Se trophic levels model. Se speciation for all trophic levels except bacteria based on bulk XANES analysis. Note that bacteria and Picocystis samples were harvested from media with an ambient concentration of 40 mg Se L⁻¹. All other samples were harvested from water with ambient concentration of approximately 9 mg Se L⁻¹.

sample was 73.5% of total Se, and protease treatment increased soluble Se recovery by 7.6 fold. The brine fly larva had no inorganic Se as measured by SAX-HPLC-ICPMS, with Se forms occurring as 4% CysSe-SeCys, 92% SeMet, and 4% SeOMet (Supporting Information Figure 4 and Table 3). The average total Se for the aquatic brine fly larva ($59.1 \pm 3.1 \mu\text{g Se}$

$\text{g}^{-1} \text{ ww}$) was elevated when compared to the shrimp and approximately 6 fold greater than in the drainage water.

Micro-XRF mapping showed that brine fly larva mostly had a uniform Se distribution throughout tissues, except for the concentrated Se hot spots associated with the intestinal tract, which is visible due to the presence of high Zn levels, while the

midgut accumulated high concentrations of Ca (Figure 2 A). Micro-XANES analysis of the localized area on the brine fly larva (Figure 2 A) showed it contained only C–Se–C forms and based on SAX-HPLC-ICPMS the C–Se–C was all in SeMet (Supporting Information Table 5).

Ephydra cinerea pupae showed accumulation of Se within the developing fly tissues with very little Se associated with the pupa cuticle, and Se was homogeneously concentrated in the pupa intestine (Figure 2B). Micro-XANES analysis of the localized areas on the brine fly pupa (Figure 2B) showed it also contained all C–Se–C forms (Supporting Information Table 5).

The total Se for the brine fly adult was highly elevated compared to all the other samples at $169 \mu\text{g Se g}^{-1} \text{ ww}$ (approximately 3 fold higher than the second highest sample brine fly larva). The highest Se concentrations in the brine fly adult were located in the thorax, with multiple Se enriched areas associated with organs in the abdomen (Figure 2 C). In the protease-XIV digested MCW extracted brine fly adult sample, Se recovery in the aqueous phase was 78%. Analysis by SAX-HPLC-ICPMS showed the brine fly adult contained mainly organic Se at 4% CysSe–SeCys, 94% SeMet, and only 2% SeO_3^{2-} (Supporting Information Table 3 and Figure 4).

In comparison, both SeMet and CysSe–SeCys were the dominant species in all insects studied in mine-impacted streams near Alberta, Canada.⁴⁶ A large scale test of Se speciation effects on invertebrate Se levels in WSJV in 1998–99 showed that aquatic animals obtain a majority of their Se directly from organic Se forms, while the inorganic forms of Se in water played a relatively minor role in animal toxicity⁴⁷ and these results agree with characterization of the Se speciation of insects living in the RRR pond.

The distribution of Se in the brine fly shows that Se is mostly accumulated in muscle tissues and intestines, which suggests Se accumulation and biotransformation is associated with either the forms of Se originally in food sources or as a result of metabolism by microbial gut microflora in the fly. Zinc was predominantly localized in larval and adult excretory organs, the Malpighian tubules (Figure 2D), as also observed in other fly such as *Drosophila*.⁴⁸ Calcium appeared to accumulate in the larval brine fly midgut and to lesser extent in the adult fly midgut (Figure 2).

Selenium Contaminated Water and Aquatic Ecosystem Management. Summary Figure 3 illustrates our findings and complete biochemical profiling of the biotransformation of Se in the different organisms in our engineered RRR drainage water aquatic ecosystem, as quantified by two different complementary analytical techniques that are contrasted in Supporting Information Table 1. Brine shrimp and brine fly at the top tropic level contained low concentrations of ($\sim 1\%$) SeO_4^{2-} or (1–2%) SeO_3^{2-} , while the predominant forms of Se accumulated were (82–94%) SeMet and (4–15%) SeCyst. Little or no trimethylselenonium was detected in brine shrimp or in brine flies, although interestingly some MeSeCys was detected in brine fly larvae. Future studies should investigate how microbial gut flora might be influencing the accumulation of these organic Se forms in these two aquatic organisms.

Selenium is very similar to sulfur in terms of its biochemical behavior in biological ecosystems. High SO_4^{2-} concentrations reduce the bioavailability of SeO_4^{2-} to a variety of organisms, including algae, bacteria, midges, daphnids, and brine shrimp.^{28,49,50} High SO_4^{2-} concentrations in the RRR pond water likely contribute to the survival of brine shrimp and brine

fly in this water. There is some evidence that invertebrates do not have efficient Se removal excretion pathways, leading to toxic levels of organic Se bioconcentration and high levels of organic Se that are in turn bioavailable to fish and birds in these environments.⁴¹

Our data elucidating all the individual forms of Se in these specific aquatic invertebrates is important because different forms have various levels of toxicity with some organic Se forms reported as nutritionally beneficial and anticarcinogenic.^{51–55} Selenomethionine, for example, has been shown to cause higher toxicity rates in birds and fish during laboratory feeding studies than observed using SeO_3^{2-} or CysSe–SeCys.^{11,56} In hypersaline ecosystems, brine flies and brine shrimp are essential components of shore birds' diet.^{45,57,58} Brine fly larvae, pupa, and adults have an even higher nutrient value per individual than brine shrimp and dominate shore bird diets at certain times of the year.^{57,59}

On the other hand, organic Se is a required micronutrient for vertebrates, including livestock and humans. A well-established fish farm industry for growing brine shrimp for harvest already exists in the United States and in Asia. An understanding of the biochemistry associated with Se-enriched brine shrimp production in order to efficiently supplement organic Se from healthful brine shrimp meals in animal diets, could lead to the development of these Se-polluted environments into sources for producing valuable bioavailable Se supplements. The closed pond system could safely provide organic Se enriched foods containing essential oils, omega 3 fatty acids, and proteins especially valuable in those areas where people are currently suffering from Se deficiency or as a Se enriched animal feed for rearing healthy livestock.

■ ASSOCIATED CONTENT

📄 Supporting Information

Details of experimental methods and additional experimental data tables and figures. This information is available free of charge via the Internet at <http://pubs.acs.org/>.

■ AUTHOR INFORMATION

Corresponding Author

*(J.L.F.) Fax: + 1-650-691-9754. E-mail: John.L.Freeman@NASA.gov. (K.R.H.) Fax: +1-414-288-7357. E-mail: krassimira.hristova@marquette.edu.

Author Contributions

○R.S., G.S.B., K.R.H., and J.L.F. contributed equally to this paper.

Notes

The authors declare no competing financial interest.

■ ACKNOWLEDGMENTS

This research was supported by a CSU Fresno California Agricultural Research Initiative granted to Dr. Gary S. Bañuelos and Dr. John L. Freeman, a Department of Energy Beam Time grant to Dr. Gary S. Bañuelos and Dr. John L. Freeman, and by the California Department of Water Resources under contract number 08-002104, task order 07-3, granted to Dr. Krassimira R. Hristova. The operations of the Advanced Light Source at Lawrence Berkeley National Laboratory were supported by the Director, Office of Science, Office of Basic Energy Sciences, U.S. Department of Energy under Contract No. DE-AC02-05CH11231. The Stanford Synchrotron Radiation Lightsource (SSRL), a Directorate of SLAC National Accelerator

Laboratory and an Office of Science User Facility is operated for the U.S. Department of Energy Office of Science by Stanford University. The SSRL Structural Molecular Biology Program is supported by the DOE, Office of Biological and Environmental Research, and by the National Institutes of Health, National Center for Research Resources, Biomedical Technology Program (P41RR001209). Ingrid J. Pickering is a Canada Research Chair and Soo In Yang was funded by the Natural Sciences and Engineering Research Council of Canada (Discovery Grant to IJP) and by a CIHR-THRUST Fellowship.

■ ABBREVIATIONS

C–Se–C	carbon-selenium-carbon
γ GMeSeCys	gamma-glutamyl-methyl-selenocysteine
MeSeCys	methyl-seleno-cysteine
μ XRF	micro-focused X-ray fluorescence
μ XANES	micro-K-edge X-ray absorption near-edge structure
bulk XANES	X-ray absorption near-edge structure
SeO ₄ ²⁻	selenate
Se ²⁻	selenide
SeO ₃ ²⁻	selenite
SeCyst	seleno-cystathionine
SeCys	seleno-cysteine
SeGSH ₂	seleno-diglutathione
SeMet	seleno-methionine
CysSe–SeCys	selenocystine
Na ₂ SeO ₄	sodium selenate
Na ₂ SeO ₃	sodium selenite
SAX-HPLC-ICPMS	strong anion exchange high performance chromatography coupled to real time inductively coupled plasma mass spectrometry
ww	wet weight
DMSe	dimethylselenide
DMDSe	dimethyldiselenide

■ REFERENCES

(1) Emmons, S. F.; Cross, W.; GH, E. Geology of the Denver basin in Colorado. *U.S. Geol. Surv. Monogr.* **1996**, *27*, 527.

(2) Deverel, S. J.; Millard, S. P. Distribution and mobility of selenium and other trace elements in shallow groundwater of the western San Joaquin Valley, California. *Environ. Sci. Technol.* **1988**, *22* (6), 697–702.

(3) Haygarth, P. M., Global importance and global cycling of selenium. In *Selenium in the Environment*; Frankenberger, W. T., Jr., Benson, S. M., Eds.; Marcel Dekker, Inc.: New York, 1994; pp 1–28.

(4) Moreno-Reyes, R.; Suetens, C.; Mathieu, F.; Begaux, F.; Zhu, D.; Rivera, M.; Boelaert, M.; Nève, J.; Perlmutter, N.; Vanderpas, J. Kashin-Beck osteoarthropathy in rural Tibet in relation to selenium and iodine status. *New Engl. J. Med.* **1998**, *339*, 1112–20.

(5) Tan, J. a.; Zhu, W.; Wang, W.; Li, R.; Hou, S.; Wang, D.; Yang, L. Selenium in soil and endemic diseases in China. *Sci. Total Environ.* **2002**, *284* (1–3), 227–235.

(6) Presser, T. S. Geologic origin and pathways of selenium from the California coast ranges to the west-central San Joaquin Valley. In *Selenium in the Environment*; Frankenberger, W. T., Jr., Benson, S. M., Eds.; Marcel Dekker, Inc.: New York, 1994; pp 139–155.

(7) Bañuelos, G.; Lin, Z. Q.; Arroyo, I.; Terry, N. Selenium volatilization in vegetated agricultural drainage sediment from the San Luis Drain, Central California. *Chemosphere* **2005**, *60* (9), 1203–1213.

(8) Bañuelos, G.; Meek, D. Accumulation of selenium in plants grown on selenium-treated soils. *J. Environ. Qual.* **1990**, *19* (4), 772–777.

(9) Bañuelos, G. Phytoremediation of selenium-contaminated soil and water produces biofortified products and new agricultural byproducts. *Dev. Uses Biofortified Agric. Prod.* **2009**, 57–70.

(10) Fan, T. W. M.; Teh, S. J.; Hinton, D. E.; Higashi, R. M. Selenium biotransformations into proteinaceous forms by foodweb organisms of selenium-laden drainage waters in California. *Aquat. Toxicol.* **2002**, *57* (1–2), 65–84.

(11) Hamilton, S. J. Review of selenium toxicity in the aquatic food chain. *Sci. Total Environ.* **2004**, *326* (1–3), 1–31.

(12) Debruyn, A. M. H.; Chapman, P. M. Selenium toxicity to invertebrates: Will proposed thresholds for toxicity to fish and birds also protect their prey? *Environ. Sci. Technol.* **2007**, *41*, 1766–1770.

(13) Ohlendorf, H. M. The birds of Kesterson Reservoir: A historical perspective. *Aquat. Toxicol.* **2002**, *57* (1–2), 1–10.

(14) Ohlendorf, H. M.; Hoffman, D. J.; Saiki, M. K.; Aldrich, T. W. Embryonic mortality and abnormalities of aquatic birds: apparent impacts of selenium from irrigation drainwater. *Sci. Total Environ.* **1986**, *52* (1–2), 49–63.

(15) Zawislanski, P. T.; Benson, S. M.; TerBerg, R.; Borglin, S. E. Selenium speciation, solubility, and mobility in land-disposed dredged sediments. *Environ. Sci. Technol.* **2003**, *37* (11), 2415–2420.

(16) Frankenberger, W.; Amrhein, C.; Fan, T.; Flaschi, D.; Glater, J.; Kartinen, E.; Kovac, K.; Lee, E.; Ohlendorf, H.; Owens, L. Advanced treatment technologies in the remediation of seleniferous drainage waters and sediments. *Irrig. Drain Syst.* **2004**, *18* (1), 19–42.

(17) Conte, F. P.; Conte, P. A. Abundance and spatial distribution of *Artemia salina* in Lake Abert, Oregon. *Hydrobiologia* **1988**, *158* (1), 167–172.

(18) Lenz, P. H.; Cooper, S. D.; Melack, J. M.; Winkler, D. W. Spatial and temporal distribution patterns of three trophic levels in a saline lake. *J. Plankton Res.* **1986**, *8* (6), 1051–1064.

(19) Bruckner, C. G.; Kroth, P. G. Protocols for the removal of bacteria from freshwater benthic diatom cultures. *J. Phycol.* **2009**, *45* (4), 981–986.

(20) Bañuelos, G. S.; Akohoue, S. Comparison of microwave digestion with block digestion for selenium and boron analysis in plant tissues. *Commun. Soil Sci. Plant Anal.* **1994**, *25* (9–10), 1655–1670.

(21) Bañuelos, G. S.; Fakra, S. C.; Walse, S. S.; Marcus, M. A.; Yang, S. I.; Pickering, I. J.; Pilon-Smits, E. A. H.; Freeman, J. L. Selenium accumulation, distribution, and speciation in spineless prickly pear cactus: A drought- and salt-tolerant, selenium-enriched nutraceutical fruit crop for biofortified foods. *Plant Physiol.* **2011**, *155* (1), 315–327.

(22) Bueno, M.; Pannier, F.; Potin-Gautier, M.; Darrouzes, J. Determination of organic and inorganic selenium species using HPLC-ICP-MS. *Agilent Technologies International* **2007**, *2007*, 1–5.

(23) Marcus, M. A.; MacDowell, A. A.; Celestre, R.; Manceau, A.; Miller, T.; Padmore, H. A.; Sublett, R. E. Beamline 10.3. 2 at ALS: a hard X-ray microprobe for environmental and materials sciences. *J. Synchrotron. Radiat.* **2004**, *11* (3), 239–247.

(24) Kelly, S.; Hesterberg, D.; Ravel, B. Analysis of soils and minerals using X-ray absorption spectroscopy. *Methods of Soil Analysis, Part 5. Mineralogical Methods*; Soil Science Society of America; Madison, WI, 2008; Vol. 5, pp 387–463.

(25) Freeman, J. L.; Zhang, L. H.; Marcus, M. A.; Fakra, S.; McGrath, S. P.; Pilon-Smits, E. A. H. Spatial imaging, speciation, and quantification of selenium in the hyperaccumulator plants *Astragalus bisulcatus* and *Stanleya pinnata*. *Plant Physiol.* **2006**, *142* (1), 124–134.

(26) Pickering, I. J.; George, G. N.; Van Fleet-Stalder, V.; Chasteen, T. G.; Prince, R. C. X-ray absorption spectroscopy of selenium-containing amino acids. *J. Biol. Inorg. Chem.* **1999**, *4* (6), 791–794.

(27) Fan, T. W. M.; Higashi, R. M.; Lane, A. N. Biotransformations of selenium oxyanion by filamentous cyanophyte-dominated mat cultured from agricultural drainage waters. *Environ. Sci. Technol.* **1998**, *32* (20), 3185–3193.

(28) Fournier, E.; Adam-Guillermin, C.; Potin-Gautier, M.; Pannier, F. Selenate bioaccumulation and toxicity in *Chlamydomonas reinhardtii*: Influence of ambient sulphate ion concentration. *Aquat. Toxicol.* **2010**, *97* (1), 51–57.

- (29) Riedel, G. F.; Sanders, J. G.; Gilmour, C. C. Uptake, transformation, and impact of selenium in freshwater phytoplankton and bacterioplankton communities. *Aquat. Microb. Ecol.* **1996**, *11*, 43–51.
- (30) Araie, H.; Shiraiwa, Y. Selenium utilization strategy by microalgae. *Molecules* **2009**, *14* (12), 4880–4891.
- (31) Wu, L. Review of 15 years of research on ecotoxicology and remediation of land contaminated by agricultural drainage sediment rich in selenium. *Ecotoxicol. Environ. Saf.* **2004**, *57* (3), 257–269.
- (32) Dungan, R.; Frankenberger, W. Microbial transformations of selenium and the bioremediation of seleniferous environments. *Bioremediat. J.* **1999**, *3* (3), 171–188.
- (33) Dungan, R. S.; Yates, S. R.; Frankenberger, W. T. Transformations of selenate and selenite by *Stenotrophomonas maltophilia* isolated from a seleniferous agricultural drainage pond sediment. *Environ. Microbiol.* **2003**, *5* (4), 287–295.
- (34) Nelson, D. C.; Casey, W. H.; Sison, J. D.; Mack, E. E.; Ahmad, A.; Pollack, J. S. Selenium uptake by sulfur-accumulating bacteria. *Geochim. Cosmochim. Acta* **1996**, *60* (18), 3531–3539.
- (35) Price, N. M.; Thompson, P. A.; Harrison, P. J. Selenium: An essential element for growth of the coastal marine diatom *Thalassiosira pseudonana* (bacillariophyceae) 1, 2. *J. Phycol.* **1987**, *23* (1), 1–9.
- (36) Harrison, P.; Yu, P.; Thompson, P.; Price, N.; Phillips, D. Survey of selenium requirements in marine phytoplankton. *Mar. Ecol.: Prog. Ser.* **1988**, *47* (1), 89–96.
- (37) Raven, J. A.; Evans, M. C. W.; Korb, R. E. The role of trace metals in photosynthetic electron transport in O₂-evolving organisms. *Photosynth. Res.* **1999**, *60* (2), 111–150.
- (38) Roesler, C. S.; Culbertson, C. W.; Etheridge, S. M.; Goericke, R.; Kiene, R. P.; Miller, L. G.; Oremland, R. S. Distribution, production, and ecophysiology of *Picocystis* strain ML in Mono Lake, California. *Limnol. Oceanogr.* **2002**, *440*–452.
- (39) Assmann, A.; Briviba, K.; Sies, H. Reduction of methionine selenoxide to selenomethionine by glutathione. *Arch. Biochem. Biophys.* **1998**, *349* (1), 201–203.
- (40) Suryo Rahmanto, A.; Davies, M. J. Catalytic activity of selenomethionine in removing amino acid, peptide, and protein hydroperoxides. *Free Radic. Biol. Med.* **2011**, *51* (12), 2288–2299.
- (41) Maier, K. J.; Nelson, C. R.; Bailey, F. C.; Klaine, S. J.; Knight, A. W. Accumulation of selenium by the aquatic biota of a watershed treated with seleniferous fertilizer. *Bull. Environ. Contam. Toxicol.* **1998**, *60* (3), 409–416.
- (42) Swift, M. C. Stream ecosystem response to, and recovery from, experimental exposure to selenium. *J. Aquat. Ecosyst. Stress Recovery* **2002**, *9* (3), 159–184.
- (43) Presser, T. S.; Piper, D. Z. Mass balance approach to selenium cycling through the San Joaquin Valley, sources to river to bay. In *Environmental Chemistry of Selenium*; Frankenberger, W.; Engberg, R. A., Eds.; Marcel Dekker: New York, 1998.
- (44) Rosetta, T. N.; Knight, A. W. Bioaccumulation of selenate, selenite, and seleno-DL-methionine by the brine fly larvae *Ephydra cinerea* Jones. *Arch. Environ. Contam. Toxicol.* **1995**, *29* (3), 351–357.
- (45) Wurtsbaugh, W. *Preliminary Analyses of Selenium Bioaccumulation in Benthic Food Webs of the Great Salt Lake, Utah*; Utah State University: Logan, UT, 2007.
- (46) Andrahennadi, R.; Wayland, M.; Pickering, I. J. Speciation of selenium in stream insects using X-ray absorption spectroscopy. *Environ. Sci. Technol.* **2007**, *41* (22), 7683–7687.
- (47) Amweg, E. L.; Stuart, D. L.; Weston, D. P. Comparative bioavailability of selenium to aquatic organisms after biological treatment of agricultural drainage water. *Aquat. Toxicol.* **2003**, *63* (1), 13–25.
- (48) Schofield, R. M.; Postlethwait, J. H.; Lefevre, H. W. MeV-ion microprobe analyses of whole *Drosophila* suggest that zinc and copper accumulation is regulated storage not deposit excretion. *J. Exp. Biol.* **1997**, *200* (24), 3235–43.
- (49) Brix, K. V.; DeForest, D. K.; Cardwell, R. D.; Adams, W. J. Derivation of a chronic site-specific water quality standard for selenium in the Great Salt Lake, Utah, U.S.A. *Environ. Toxicol. Chem.* **2004**, *23* (3), 606–612.
- (50) Forsythe, B. L.; Klaine, S. J. The interaction of sulfate and selenate (Se₆₊) effects on brine shrimp, *Artemia* I spp. *Chemosphere* **1994**, *29* (4), 789–800.
- (51) Nogueira, C. W.; Rocha, J. o. B. Toxicology and pharmacology of selenium: Emphasis on synthetic organoselenium compounds. *Arch. Toxicol.* **2011**, *85* (11), 1313–1359.
- (52) Clark, L.; Combs, G., Jr; Turnbull, B.; Slate, E.; Alberts, D.; Abele, D.; Allison, R.; Bradshaw, J.; Chalker, D.; Chow, J. The nutritional prevention of cancer with selenium 1983–1993: A randomized clinical trial. *JAMA* **1996**, *276*, 1957–63.
- (53) Finley, J. W.; Davis, C. D.; Feng, Y. Selenium from high selenium broccoli protects rats from colon cancer. *J. Nutrition* **2000**, *130* (9), 2384–2389.
- (54) Ellis, D. R.; Salt, D. E. Plants, selenium and human health. *Curr. Opin. Plant Biol.* **2003**, *6* (3), 273–279.
- (55) Ip, C.; Ganther, H. Relationship between the chemical form of selenium and anticarcinogenic activity. *Cancer Chemoprev.* **1992**, 479–488.
- (56) Heinz, G. H.; Hoffman, D. J.; Gold, L. G. Impaired reproduction of mallards fed an organic form of selenium. *J. Wildl. Manage.* **1989**, 418–428.
- (57) Herbst, D. B. Salinity controls on trophic interactions among invertebrates and algae of solar evaporation ponds in the Mojave Desert and relation to shorebird foraging and selenium risk. *Wetlands* **2006**, *26* (2), 475–485.
- (58) Wrege, P. H.; Hite, J. M.; Winkler, D. W. *The Diet of California Gull Nestlings at Mono Lake: Seasonal and Diurnal Variation*; Point Reyes Bird Observatory: Petaluma, CA, 2001.
- (59) Collins, N. Population ecology of *Ephydra cinerea* Jones (Diptera: Ephydriidae), the only benthic metazoan of the Great Salt Lake, U.S.A. *Hydrobiologia* **1980**, *68* (2), 99–112.